



Diurnal Fluctuations in Acidification and Hypoxia Reduce Growth and Survival of Larval and Juvenile Bay Scallops (*Argopecten irradians*) and Hard Clams (*Mercenaria mercenaria*)

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OPEN ACCESS

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Specialty section:

This article was submitted to Marine Pollution, a section of the journal Frontiers in Marine Science

Received: 30 September 2016 Accepted: 14 December 2016 Published: 04 January 2017

Citation:

Gobler CJ, Clark HR, Griffith AW and Lusty MW (2017) Diurnal Fluctuations in Acidification and Hypoxia Reduce Growth and Survival of Larval and Juvenile Bay Scallops (Argopecten irradians) and Hard Clams (Mercenaria mercenaria). Front. Mar. Sci. 3:282. doi: 10.3389/fmars.2016.00282

Diurnal variations in pH and dissolved oxygen (DO) concentrations are common seasonal phenomena in many eutrophic estuaries, yet few studies have investigated the concurrent effects of low pH and low DO on marine organisms inhabiting these coastal systems. Here, we assess the effects of diurnal variations in pH and DO on the early-life history of two bivalve species native to Northeast US estuaries, the bay scallop (Argopecten irradians) and hard clam (Mercenaria mercenaria). In one set of experiments, larval- and juvenile-life stage bivalves were exposed to ambient conditions (pH_T \sim 7.9), two continuously-low pH levels (pH_T \sim 7.3 and 7.6), and diurnal fluctuations between the ambient and low conditions yielding mean pH levels equal to the intermediate pH levels. In a second set of experiments, larval bivalves were exposed to ambient conditions $(pH_T \sim 7.9, DO \sim 7 \text{ mg L}^{-1})$, two levels of low pH and DO $(pH_T \sim 7.2, DO \sim 1 \text{ mg L}^{-1})$; $pH_T \sim 7.4$, DO $\sim 4 \text{ mg L}^{-1}$) and diurnal fluctuations of both pH and DO between the ambient and low pH/DO levels that resulted in mean pH and DO levels equal to the intermediate pH and DO levels. Diurnal acidification treatments with ambient DO levels vielded survival rates for both species at both life stages that were consistent with the survival of individuals exposed to the same mean level of chronic pH with juveniles being more resistant to acidification than larvae. In contrast, when both pH and DO varied diurnally, the survival rates of larval bivalves were significantly lower than the survival of individuals chronically exposed to the same mean levels of pH and DO, an indication that bivalves were physiologically more vulnerable to concurrent fluctuations of both parameters compared to acidification alone. While both species displayed sensitivities to diurnal fluctuations in pH and DO, scallops were relatively more susceptible than hard clams. Since many shallow eutrophic estuaries presently experience diurnal cycles of both pH and DO when early-life stages of bivalves are present in estuaries, the populations of the bivalves studied are likely impacted by these conditions which are likely to intensify with climate change.

Keywords: ocean acidification, hypoxia, climate change, Argopecten irradians, Mercenaria mercenaria

INTRODUCTION

Two consequences of climate change are acidification and deoxygenation of world oceans (Doney et al., 2012). Beyond anthropogenically-induced changes in climate, acidification, and low oxygen conditions are inextricably linked in the ocean via the processes of respiration and photosynthesis and often display highly similar patterns in marine ecosystems (Feely et al., 2010; Cai et al., 2011; Wallace et al., 2014). In temperate coastal zones, hypoxia, or low oxygen conditions as well as acidification develop seasonally as seawater warms, bottom waters become isolated from surface waters via stratification, and rates of respiration accelerate. Many estuaries in particular, are net-heterotrophic due strong internal and external sources of organic carbon, and thus on a net-annual basis produce excess CO₂ and consume O₂ (Caffrey, 2004; Del Giorgio and Williams, 2005). In urbanized regions receiving excessive nutrient loads (Melzner et al., 2012; Wallace et al., 2014) or within eutrophied river plumes (Cai et al., 2011) hypoxia and acidification can occur at extreme levels. Beyond the role of excessive nutrient loads in driving this trend, specific estuarine habitats such as salt marshes are naturally enriched in organic carbon that when respired can create hypoxic and acidified conditions, particularly within warmer waters (Ringwood and Keppler, 2002; Baumann et al., 2014).

The persistence of hypoxia and acidification in coastal zones can vary from hours (Baumann et al., 2014) to months (Wallace et al., 2014), depending on the intensity of respiration and the hydrodynamics of ecosystems. In shallow, well-mixed estuaries with high rates of respiration, hypoxia and acidification can occur diurnally, as photosynthesis during the day results in high DO and pH levels, while respiration during night decrease DO and pH levels. The intensity of this process can be related to the depth of the water column, given that sediments are known to be strong sources of CO₂ and sinks of O₂ (Green and Aller, 1998) and their influence is inversely proportional to the depth of the water column (Gobler and Baumann, 2016). Within deeper, stratified ecosystems including oxygen minimum zones (OMZs), hypoxic and acidified water can persist for weeks, months, or longer, as benthic and/or deep water respiration continually consumes oxygen and produces CO₂ faster than it is replenished via diffusion and mixing from surface waters (Paulmier et al., 2011; Wallace et al., 2014).

Research regarding of how concurrent acidification and hypoxia impacts marine life is in its infancy. For decades, studies have documented how low oxygen affects ocean organisms, but these studies either ignored pH levels or, in most cases, treated seawater with nitrogen gas and thus basified, rather than acidified, seawater (Gobler et al., 2014). There have been few studies that have examined how concurrent low oxygen and low pH affect marine life. Some studies have found that low oxygen effects tend to dominate over pH effects (Kim et al., 2013) while others have found that the combined effects of acidification and hypoxia are more severe than the effects of each individual stressors and that impacts in bivalves can be age-dependent (Gobler et al., 2014; Steckbauer et al., 2015). Keppel et al. (2016) reported that diel-cycling of DO decreased growth rates of the eastern oyster, *Crassostrea virginica*, but diel-cycling of pH did not. Finally, Clark and Gobler (2016) compared the effects of chronic hypoxia and acidification to diurnal fluctuations in CO_2 and O_2 on larval bivalves and found that diurnal exposure to low pH and/or low DO rarely altered, and never fully eliminated, the negative effects of hypoxia and/or acidification even when the diurnal cycles yielded higher mean pH and DO levels. This outcome suggested that the magnitude of pH and DO fluctuations may be too intense and/or the duration of normoxic and normocapnic conditions are not of a long enough duration to permit the bivalve larvae to overcome the physiological stress of hypoxia and acidification.

The purpose of this study was to expand on the work conducted by Clark and Gobler (2016) and to further clarify the effects of static vs. diurnally-fluctuating low DO and low pH on early-life stage bivalves. First, this study compared responses of larval and juvenile life-staged bay scallops (*Argopecten irradians*) and hard clams (*Mercenaria mercenaria*) to chronically-low pH and diurnal fluctuations in pH. Next, this study contrasted diurnal acidification only with that of diurnal changes in pH and DO in unison, a condition that occurs naturally in ecosystems (Baumann et al., 2014; Wallace et al., 2014). Finally, this study established two separate levels of chronically-low pH and DO to compare outcomes to those found for individuals exposed to the diurnal fluctuations to understand whether responses were more similar to the mean or minimum exposure levels of pH and DO.

METHODS

Manipulation of Carbonate Chemistry and DO

Replicate (n = 4) 8 L polyethylene vessels were used for experiments and filled with UV-sterilized, 0.2 µm filtered seawater from Old Fort Pond in Shinnecock Bay, NY, USA (salinity = 30). A constant temperature of $23-24^{\circ}C$ was maintained by placing experimental vessels in water baths heated by Delta Star[®] heat pumps. Two types of experiments were conducted. The first involved the manipulation of only pH while the second involved simultaneous manipulations of both pH and DO. For first set of experiments, treatments of control (\sim 7.9), intermediate (\sim 7.5), and low pH_T (\sim 7.2) were established and carbonate chemistry was maintained by bubbling mixtures of concentrated 5% CO2 gas and ambient air into the experimental vessels (details below). An additional treatment of diurnal pH was included where pH oscillated between control and low pH conditions every 12 h (details below; Figure 1). For the second set of experiments, pH and DO levels were altered in unison and four treatments were established: A control treatment (pH_T \sim 7.9, DO \sim 7.0 mg L⁻¹), a chronic-intermediate pH-DO treatment (~7.5, ~4.0 mg L⁻¹), a chronic-low pH-DO treatment (~7.2, \sim 2.0 mg L⁻¹), and a diurnally-fluctuating pH-DO treatment that resulted in the control (e.g., ambient conditions) pH and DO levels by day and the low pH and DO levels at night. Daily-mean levels were similar to the intermediate pH-DO treatments (~7.5, ~4.0 mg L^{-1} ; details below; Figure 1). The range of pH and DO variations in all diurnal treatments was consistent with daily fluctuations observed within temperate estuaries in the Northeast



US (Ringwood and Keppler, 2002; Baumann et al., 2014; Wallace et al., 2014).

The delivery rate of gases was controlled by a series of Cole-Parmer[®] gas regulators, single-channel flowmeters, and/or multi-channel gas proportioners. Carbon dioxide gas was used to manipulate pH levels while and nitrogen gas to manipulate DO concentrations as described by Gobler et al. (2014). To mimic diurnal patterns of pH and DO concentrations, Alcon solenoid valves were attached to compressed gas tanks and ambient air lines and were controlled with a Rain Bird-[®] timer (Clark and Gobler, 2016). During daytime cycles (0900–2100 h), valves on the ambient airlines were opened and the valves controlling the flow for mixed gases (CO₂, N₂, and CO₂N₂) were closed to establish ideal pH and DO (pH_T = 7.9; DO = 7.0 mg L⁻¹) conditions. At night (2100–0900 h), the valves on the appropriate CO₂, N₂, or CO₂N₂ gas tanks were opened to establish low pH and/or DO conditions.

Measuring Salinity, Temperature, pH, and DO

Salinity was measured using a YSI 600QS multi-parameter water quality sonde and temperature was logged every 15 min using a HOBO[®] U-002-64 Data Logger (Onset). Measurements of pH_T were made daily using a Honeywell Durafet Ion Sensitive Field Effect Transistor (ISFET)-based pH sensor calibrated with a seawater pH standard (Dickson, 1993). Every 15 min, in the diurnal treatments, pH values were measured and logged with a Thermo-Scientific Orion STAR A321 pH meter to quantify diurnal patterns of pH. A Clark-type electrode YSI 5100 oxygen meter was used to make daily DO measurements and DO was logged every 15 min on a HOBO[®] U26 dissolved oxygen logger (Onset) in the diurnal treatments. Prior studies have found these instruments measure levels of dissolved oxygen that are indistinguishable from discrete measurements made with Winkler titrations (Gobler et al., 2014).

Samples were obtained for dissolved inorganic carbon (DIC) analysis at the beginning and end of each experiment and were quantified using an EGM-4, Environmental Gas Analyzer-® (PP Systems) system after separation of gas and liquid phases with a Liqui-Cel-® Membrane (Membrana). For all diurnal treatments, samples were collected and analyzed from the end of both the day and night cycles. To determine the precision and accuracy of DIC measurements, Dr. Andrew Dickson's (University of California San Diego, Scripps Institution of Oceanography) certified reference material for DIC was analyzed during each analytical run (mean percent recovery of DIC across all analytical runs: $103 \pm 6\%$). DIC levels, along with pH, temperature, salinity, pressure, phosphate, silicate, and carbonic acid dissociation constants recommended for estuarine waters (Millero et al., 2006) were analyzed with the CO2SYS program (http://cdiac. ornl.gov/ftp/co2sys/) in order to quantify levels of pCO_2 , $\Omega_{calcite}$, $\Omega_{aragonite}$ carbonate, and total alkalinity. Since pH values in some experimental treatments fluctuated widely each day and since pH is on a log scale, mean pH values were determined by first converting pH to $[H^+]$ concentrations. Mean $[H^+]$ concentrations were then converted back to pH.

Care and Maintenance of Organisms

All organisms used for this study were marine invertebrates and as such, ethical approval for their use was not required. Larval and juvenile A. irradians and M. mercenaria were obtained from the East Hampton Town Shellfish Hatchery located in Montauk, NY, USA. Experimental vessels (8 L) were stocked with either 10,000 D-stage larvae (<24 h) or 15 juveniles (initial size of juvenile A. irradians = 2 ± 0.1 mm; M. mercenaria = 7 \pm 0.5 mm). All shellfish were fed a diet of live-cultured algae (4 \times 10⁴ cells mL⁻¹; *Isochrysis galbana*) daily (Kraeuter and Castagna, 2001; Helm et al., 2004). Full water changes were performed twice-weekly for experiments involving larval shellfish, during which the entire contents of each experimental vessel was carefully poured over a 64 µm sieve to capture larvae. Larvae collected on the sieve were condensed into a 50 mL container and 2 mL aliquots were removed and preserved with a 3% solution of buffered formalin phosphate to assess mortality, size (distance from the umbo to furthest-leading ventral edge), and developmental stage (veliger, pediveliger, or fully-metamorphosed) at each time-point using a compound microscope with Nikon DigiSight Color Digital Camera System (DSVi1) and ImageJ[©] software. Live larvae were distinguishable from dead larvae by the pigmentation, presence or absence of internal organs, and whether or not valves were fully intact. Percent metamorphosis was determined based upon the total number of surviving larvae at each time-point. Larval experiments concluded when the majority of individuals had fully metamorphosed within control treatments, 11 and 15 days for M. mercenaria and A. irradians respectively. Final growth rates (μ m day⁻¹) were calculated using initial (length at 24 h post-fertilization) and final shell lengths.

For experiments involving juvenile shellfish, water was exchanged in each experimental vessel twice weekly.

Survival was assessed daily, whereby dead individuals, deemed so by their gaping shells and/or lack of response to external stimulate, were removed. Initial and final size, determined using digital calipers, was used to determine growth rates. Juvenile experiments were concluded after ~ 1 month.

Data Analysis

All statistical analyses were performed using RStudio[©] (http:// www.cran.r-project.org) statistical software. Survival and development data were arcsine-square-root transformed before analysis. A one-way analysis of variance (ANOVA) was used to detect differences in survival, development, and growth among treatments for all experiments (**Table 1**). A Shapiro-Wilk's test was used to verify experimental data was normally distributed and Bartlett's test was used confirm homogeneity of variance amongst each dataset. When significant differences were detected, a Tukey's honest significant difference test (Tukey HSD) was used to identify the source of variance. All results were deemed significant at $\alpha \leq 0.05$.

RESULTS

Diurnal Acidification Experiments with Larval Bivalves

Survival of larval *A. irradians* was significantly reduced by low pH (one-way ANOVA; p < 0.001; **Figure 2A**; **Table 1**). Larvae exposed to control and intermediate pH levels (pH_T = 7.94 ± 0.06 and 7.60 ± 0.03, respectively) experienced 36 ± 2 and 21 ± 11% survival (±standard deviation), respectively, whereas the percentage of larvae that survived in low pH treatments (pH_T = 7.33 ± 0.07; survival = 7 ± 5%) was significantly lower than the intermediate (Tukey HSD; p < 0.05) and control (Tukey HSD; p < 0.01) treatments, but not different than the diurnal treatments (mean pH_T = 7.47 ± 0.43; survival = 15 ± 3%; **Figure 2A**; **Table 1**). Low pH conditions slowed larval growth (one-way ANOVA; p < 0.001) relative to control treatments, whereas intermediate and diurnal pH treatments had no discernible effect

on growth or development (**Figures 2B,C**). Larvae grew at a rate of $11 \pm 5 \ \mu m \ day^{-1}$ in low pH treatments, $26 \pm 4 \ \mu m \ day^{-1}$ in control treatments, $20 \pm 2 \ \mu m \ day^{-1}$ in intermediate pH treatments, and $22 \pm 3 \ \mu m \ day^{-1}$ in diurnal pH treatments (**Figure 2B**). Twelve days post-fertilization, 88 ± 5 , 58 ± 22 , and $61 \pm 19\%$ of *A. irradians* larvae had metamorphosed in the control, intermediate, and diurnal treatments respectively, while only $13 \pm 19\%$ had metamorphosed in low pH treatments (**Figure 2C**).

Low pH (pH_T = 7.29 ± 0.06) significantly reduced survival (one-way ANOVA; p < 0.05) of larval *M. mercenaria*, although there was no effect of intermediate (pH_T = 7.58 ± 0.04) or diurnally-fluctuating pH treatments (mean pH_T = 7.54 ± 0.36; **Figure 3A**; **Table 2**). Larvae reared under control (pH_T = 7.91 ± 0.02), intermediate, and diurnal pH conditions displayed survival rates of 27 ± 3, 23 ± 2, and 25 ± 5%, respectively, whereas the survival of larvae in the low pH treatment was 15 ± 5%, significantly lower than all other treatments (Tukey HSD; p < 0.05; **Figure 3A**; **Table 2**). Growth and development of *M. mercenaria* larvae did not differ significantly among experimental treatments (one-way ANOVA; p > 0.05; **Figures 3B,C**).

Diurnal Acidification Experiments with Juvenile Bivalves

Relative to control (e.g., ideal) pH treatments, survival of juvenile *A. irradians* was significantly reduced by low pH (pH_T = 7.13 \pm 0.03; Tukey HSD; p < 0.05), intermediate pH (pH_T = 7.49 \pm 0.04; Tukey HSD; p < 0.01), and diurnally-fluctuating pH conditions (mean pH_T = 7.57 \pm 0.43; Tukey HSD; p < 0.001; **Figure 4A**; **Table 3**). In the control treatment (pH_T = 7.92 \pm 0.05), 92 \pm 6% of individuals survived whereas low, intermediate, and diurnally-fluctuating pH treatments displayed 57 \pm 25, 47 \pm 9, and 40 \pm 12% survival, respectively, levels significantly lower than the control (Tukey HSD; p > 0.05; **Figure 4A**). There was no effect of pH on the growth of juvenile *A. irradians* (**Figure 4B**). The differing levels of pH (**Table 4**) used in experiments did not significantly alter the survival and growth of juvenile *M. mercenaria* (**Figures 5A,B**).

TABLE 1 Mean (± standard deviation) pH, pCO2, saturation states of calcite and aragonite, total dissolved inorganic carbon (TDIC), carbonate, total
alkalinity (TA), salinity, and temperature for the larval Argopecten irradians diurnal acidification experiment.

Parameter		Continuous			Diurnal		
	Control	Intermediate	Low	Day	Night	Mean	
pH _T	7.94 ± 0.06	7.60 ± 0.03	7.33 ± 0.07	7.87 ± 0.02	7.32 ± 0.08	7.47 ± 0.43	
pCO ₂ (µatm)	516 ± 18	1200 ± 134	2340 ± 384	662 ± 148	2700 ± 907	1680 ± 527	
$\Omega_{calcite}$	3.44 ± 0.82	1.61 ± 0.10	0.90 ± 0.16	3.15 ± 0.69	0.99 ± 0.20	2.07 ± 0.45	
$\Omega_{aragonite}$	2.21 ± 0.53	1.04 ± 0.06	0.58 ± 0.10	2.03 ± 0.44	0.64 ± 0.13	$1.33 \pm .03$	
TDIC (μ mol L ⁻¹)	1930 ± 269	1970 ± 94	2030 ± 82	2090 ± 473	2290 ± 467	2190 ± 470	
CO_3^{2-} (µmol L ⁻¹)	136 ± 33	63.7 ± 3.9	35.4 ± 6.1	125 ± 28	39.2 ± 7.9	81.9 ± 18	
TA (μ mol L ⁻¹)	2110 ± 307	2030 ± 89	2010 ± 80	2250 ± 495	2260 ± 444	2250 ± 469	
Salinity	28.3 ± 0.58	28.3 ± 0.58	28.3 ± 0.58	28.3 ± 0.58	28.3 ± 0.58	28.3 ± 0.58	
Temperature (°C)	23.5 ± 0.52	23.5 ± 0.52	23.5 ± 0.52	23.5 ± 0.52	23.5 ± 0.52	23.5 ± 0.52	



Diurnal Acidification and Hypoxia Experiments with Larval Bivalves

Chronic and diurnal exposure to low pH and low DO significantly (one-way ANOVA; p < 0.05) reduced survival of larval A. irradians (Figure 6A). The survival of larval A. irradians in the control (38 \pm 11%; pH_T = 7.89 \pm 0.00; DO = 6.87 \pm 0.25 mg L⁻¹) was significantly (Tukey HSD; p < 0.05) greater than larvae at intermediate levels ($26 \pm 6\%$; pH_T = 7.48 ± 0.05 ; $DO = 4.08 \pm 0.41 \text{ mg L}^{-1}$) which was significantly (Tukey HSD; p < 0.05) greater than larvae exposed to low (16 \pm 4%; pH_T = 7.22 \pm 0.05; DO = 1.38 \pm 0.45 mg L⁻¹) and diurnal (6 \pm 1%; mean pH_T = 7.61 \pm 0.26; mean DO = 4.11 \pm 2.80 mg L^{-1}) treatments (**Figure 6A**; **Table 5**). Both the low and diurnal treatment slowed growth (one-way ANOVA; p < 0.001) from $13 \pm 1 \,\mu\text{m} \text{ day}^{-1}$ in both the control and intermediate pH-DO treatments to 7 \pm 1 and 10 \pm 0.4 μm day $^{-1}$ in the low and diurnal pH-DO treatments (Figure 6B). Continuously-low pH and DO also significantly delayed development (one-way ANOVA; p <



0.05), while exposure to intermediate and diurnally-fluctuating pH and DO had no effect. After 15 days, 32 ± 8 , 36 ± 4 , and $32 \pm 13\%$ of larvae had metamorphosed in the control, intermediate, and diurnal pH-DO treatments respectively, whereas only 13 \pm 3% had metamorphosed in the low pH-DO treatment (**Figure 6C**).

Exposure of larval *M. mercenaria* to low (pH_T = 7.24 ± 0.04; DO = 1.32 ± 0.30 mg L⁻¹) and diurnally-fluctuating pH-DO (mean pH_T = 7.41 ± 0.34; mean DO = 4.02 ± 3.00 mg L⁻¹) significantly (Tukey HSD; p < 0.05) reduced survival to 15 ± 5 and 10 ± 4% compared to the control (pH_T = 7.87 ± 0.03; DO = 6.92 ± 0.13 mg L⁻¹) and intermediate (pH_T = 7.43 ± 0.03; DO = 3.92 ± 0.34 mg L⁻¹) pH treatments that had survival of 29 ± 4 and 20 ± 4%, respectively (**Figure 7A; Table 6**). Continuously-low pH and DO conditions significantly (p < 0.05; Tukey HSD) slowed the growth of *M. mercenaria* larvae relative to control treatments. Growth among the remaining

Parameter	Continuous			Diurnal		
	Control	Intermediate	Low	Day	Night	Mean
pH _T	7.91 ± 0.02	7.58 ± 0.04	7.29 ± 0.06	7.82 ± 0.06	7.27 ± 0.05	7.54 ± 0.34
pCO ₂ (µatm)	643 ± 169	1390 ± 309	2800 ± 593	792 ± 320	2900 ± 710	1850 ± 515
$\Omega_{calcite}$	2.53 ± 0.09	1.24 ± 0.07	0.67 ± 0.08	1.99 ± 0.21	0.61 ± 0.07	1.30 ± 0.14
$\Omega_{aragonite}$	1.63 ± 0.06	0.80 ± 0.05	0.43 ± 0.06	1.28 ± 0.13	0.40 ± 0.05	0.84 ± 0.09
TDIC (μ mol L ⁻¹)	1820 ± 238	1850 ± 205	1940 ± 253	1750 ± 294	1900 ± 269	1820 ± 282
CO_3^{2-} (µmol L ⁻¹)	100 ± 3.9	49.2 ± 3.2	26.5 ± 3.7	78.9 ± 7.7	24.4 ± 3.0	51.6 ± 5.4
TA (μ mol L ⁻¹)	1950 ± 234	1880 ± 197	1890 ± 240	1850 ± 278	1850 ± 251	1850 ± 264
Salinity	28.5 ± 3.5					
Temperature (°C)	23.0 ± 0.25	23.0 ± 0.25	23.0 ± 0.25	23.0 ± 0.25	23.0 ± 0.25	23.0 ± 0.25

TABLE 2 | Mean (± standard deviation) pH, pCO₂, saturation states of calcite and aragonite, total dissolved inorganic carbon (TDIC), carbonate, total alkalinity (TA), salinity, and temperature for the larval *Mercenaria mercenaria* diurnal acidification experiment.



treatments was not statistically different (p > 0.05; Tukey HSD). Delays in development were only observed in *M. mercenaria* larvae exposed to lower pH and DO (Tukey HSD); p < 0.05; **Figure 7C**). Eleven days post-fertilization, $61 \pm 4\%$ of larvae had metamorphosed in the control treatment, whereas only 41 ± 7 , 28 ± 9 , and $24 \pm 7\%$ had metamorphosed in the intermediate (Tukey HSD; p < 0.05), diurnal (p < 0.001), and low pH-DO (Tukey HSD; p < 0.001) treatments respectively (**Figure 7C**).

DISCUSSION

Low oxygen and low pH conditions are a common feature of many eutrophic estuaries today and these conditions are expected to be intensified by climate change in coming decades (Diaz and Rosenberg, 2008; Doney et al., 2009). While the detrimental effects of hypoxia and acidification on marine life have been well-studied, the combined impacts of these stressors are poorly understood (Gobler et al., 2014). Early-life stage bivalves are spawned during early-summer months, a period when pH and DO can be concurrently low but also undergo sharp diurnal changes (Ringwood and Keppler, 2002; Wootton et al., 2008; Baumann et al., 2014). Prior studies investigating the effects of continuous acidification and hypoxia on early-life stage bivalves have found these stressors can have additive and synergisticallynegative effects (Gobler et al., 2014) and that diurnal variation in these conditions do not provide a refuge for larval-stage bivalves (Clark and Gobler, 2016). This study builds on these findings by demonstrating that although diurnal exposure of larvae to low pH alone yielded survival rates consistent with chronic exposure to similar mean pH conditions, the concurrent cycling of both pH and DO to low levels on diurnal time-scales yielded lower survival outcomes than individuals exposed to chronic levels of the same mean levels (e.g., intermediate treatments) of pH and DO. This study further demonstrated that juvenile-stage bivalves are more resistant to acidification, both chronic and diurnal, than larval stages and that A. irradians is more vulnerable to acidification than M. mercenaria. Collectively, these findings bring novel insight regarding the fate of bivalve populations in current and future coastal ecosystems.

During the past decade, acidification has been identified as a considerable environmental threat to early-life stage bivalves (Talmage and Gobler, 2009; Gazeau et al., 2013; Kroeker et al., 2013). Early-shell formation in bivalve larvae is an energetically expensive process that becomes an even greater challenge when acidification reduces the availability of carbonate ions (Pörtner, 2008; Waldbusser et al., 2013) and recent studies have emphasized the significant physiological stress wrought by enhanced acid-base regulation (Pörtner et al., 2004; Sokolova, 2013; Waldbusser et al., 2015). While chronic exposure to

Parameter	Continuous			Diurnal		
	Control	Intermediate	Low	Day	Night	Mean
pH _T	7.92 ± 0.05	7.49 ± 0.04	7.13 ± 0.03	7.87 ± 0.05	7.51 ± 0.10	7.57 ± 0.43
pCO ₂ (µatm)	527 ± 132	1510 ± 146	3070 ± 375	990 ± 232	522 ± 50	756 ± 141
$\Omega_{calcite}$	3.32 ± 0.18	1.34 ± 0.15	0.52 ± 0.10	2.35 ± 0.41	3.15 ± 0.23	2.75 ± 0.32
$\Omega_{aragonite}$	2.16 ± 0.12	0.88 ± 0.09	0.34 ± 0.06	1.53 ± 0.27	2.05 ± 0.15	1.79 ± 0.21
TDIC (μ mol L ⁻¹)	1860 ± 259	1950 ± 96	1760 ± 235	2090 ± 306	1810 ± 26	1950 ± 166
CO_3^{2-} (µmol L ⁻¹)	134 ± 7.4	54.2 ± 5.8	21.2 ± 3.9	94.6 ± 17	127 ± 9.3	111 ± 13
TA (μ mol L ⁻¹)	2050 ± 254	1990 ± 98	1700 ± 230	2200 ± 305	1990 ± 21	2090 ± 163
Salinity	30.7 ± 0.58	30.7 ± 0.58				
Temperature (°C)	23.7 ± 0.48	23.7 ± 0.48	23.7 ± 0.48	23.7 ± 0.48	23.7 ± 0.48	23.7 ± 0.48

TABLE 3 | Mean (\pm standard deviation) pH, pCO_2 , saturation states of calcite and aragonite, total dissolved inorganic carbon (TDIC), carbonate, total alkalinity (TA), salinity, and temperature for the juvenile *Argopecten irradians* diurnal acidification experiment.

TABLE 4 | Mean (\pm standard deviation) pH, pCO_2 , saturation states of calcite and aragonite, total dissolved inorganic carbon (TDIC), carbonate, total alkalinity (TA), salinity, and temperature for the juvenile *Mercenaria mercenaria* diurnal acidification experiment.

Parameter	Continuous			Diurnal		
	Control	Intermediate	Low	Day	Night	Mean
pH _T	7.86 ± 0.05	7.57 ± 0.03	7.25 ± 0.03	7.83 ± 0.05	7.35 ± 0.12	7.57 ± 0.33
pCO ₂ (µatm)	561 ± 93	1140 ± 65	2500 ± 267	605 ± 112	1940 ± 391	1270 ± 252
$\Omega_{calcite}$	2.52 ± 0.16	1.38 ± 0.16	0.69 ± 0.09	2.30 ± 0.30	0.88 ± 0.32	1.59 ± 0.31
$\Omega_{aragonite}$	1.62 ± 0.10	0.89 ± 0.12	0.44 ± 0.06	1.48 ± 0.20	0.57 ± 0.21	1.02 ± 0.21
TDIC (μ mol L ⁻¹)	1690 ± 117	1760 ± 107	1850 ± 155	1670 ± 213	1770 ± 136	1720 ± 174
CO_3^{2-} (µmol L ⁻¹)	98.9 ± 6.3	54.3 ± 7.8	27.1 ± 3.7	90.6 ± 13	34.6 ± 13	62.6 ± 13
TA (μ mol L ⁻¹)	1820 ± 116	1800 ± 117	1810 ± 154	1790 ± 222	1760 ± 166	1780 ± 194
Salinity	27.5 ± 2.1					
Temperature (°C)	23.3 ± 0.61	23.3 ± 0.61	23.3 ± 0.61	23.3 ± 0.61	23.3 ± 0.61	23.3 ± 0.61

acidification has been shown to lead to enhanced mortality in larval and juvenile *A. irradians*, larval *M. mercenaria*, and slowed growth in juvenile *M. mercenaria* (Talmage and Gobler, 2010, 2011; Gobler and Talmage, 2013), in this study exposure to only diurnal acidification provided some relief from these effects. For example, for both species, larval-stage growth, survival and development were not statistically different to the rates found within the chronic-intermediate acidification treatments with highly similar mean pH levels, despite nightly excursions into extremely low pH conditions. This finding implies that when experiencing only the stress of acidification, these bivalve larvae can resist, daily extreme acidification (pH_T ~ 7.3) if such excursions are followed by extended normal pH conditions.

This study documented stage-specific vulnerability of bivalves to acidification. The larval life stages of both species were highly vulnerable to low pH conditions in terms of growth, survival, and development, while juvenile-life stages were less affected. Neither species exhibited slower growth in response to low pH conditions, and only bay scallops displayed reduced survivorship. Prior work has demonstrated that bivalve larvae tend to be the most sensitive life stage to acidification (Widdicombe and Spicer, 2008; Talmage and Gobler, 2011; Gazeau et al., 2013), likely due to the very large energy expenditure associated with initial calcium carbonate shell formation (Waldbusser et al., 2013, 2014). For example, within 48 h of fertilization, larvae precipitate roughly 90% of their body weight as calcium carbonate (Waldbusser et al., 2013), a process that is severely disrupted under low pH conditions (Gobler and Talmage, 2013; Waldbusser et al., 2013).

The larval-life stages can be a significant population bottleneck for bivalve molluscs (Kraeuter et al., 1981; Kraeuter and Castagna, 2001). In the field, <1% of spawned individuals reach post-set (i.e., juvenile) life-stages (Vance, 1973; Fegley, 2001). Here, we demonstrate that extreme-diurnal variation in pH and DO in coastal areas may further limit shellfish recruitment in highly productive or eutrophic regions experiencing such conditions. Future climate changes may depress the baseline and/or enhance the intensity of such diurnal fluctuations (Wootton et al., 2008; Miller et al., 2009) and may add additional selective pressure on shellfish populations, especially during initial life stages. Prior studies have shown that acidification induced mortality can depress the productivity of whole bivalve populations (O'Leary et al. submitted). Further, bivalve shellfish perform several ecological functions within coastal marine ecosystems such as the establishment of critical



habitat (Tolley and Volety, 2005; Abeels et al., 2012), decreasing coastal erosion (Meyer et al., 1997; Coen et al., 2007), and improve the surrounding water quality through the removal of suspended particulates (Officer et al., 1982; Wall et al., 2008). Finally, shellfish also provide a significant source of income and economic stimulus within coastal communities (Shumway et al., 2003; Ekstrom et al., 2015). Therefore, extreme variations in pH and DO on diurnal time-scales may represent a current and potentially growing threat to these ecologically and economically valuable marine resources.

Juvenile stages of bivalves were less vulnerable to acidification than larvae stages. In contrast to larvae, juvenile-stage bivalves can rely on internal stores of inorganic carbon to synthesize calcium carbonate shell (Gazeau et al., 2013), making them less vulnerable to physiological stressors. For hard clams, *M. mercenaria*, the age-specific vulnerability has been identified previously, with older individuals such as those used in this study being more resistant to low pH conditions (Green et al., 2009; Gobler et al., 2014). Interestingly, such resistance can be lost in juvenile hard clams when they are concurrently exposed to acidification and hypoxia (Gobler et al., 2014), a finding consistent with the results of this study, at least for larval stages (see below).



During this study, both stages of M. mercenaria were markedly more tolerant to acidification than A. irradians, a finding consistent with prior studies of these two species (Talmage and Gobler, 2009, 2010). Juvenile hard clams were entirely tolerant of the low pH they were exposed to, a trait potentially related to their position in the seabed relative to bay scallops. Scallops are epifaunal, live on the sediment surface, and are thus exposed to overlying water column chemistry, whereas hard clams are infaunal, burrowed in coastal sediments that are naturally acidified and are often exposed to porewaters that are undersaturated in calcium carbonate (Green and Aller, 1998; Green et al., 2009). The differential vulnerability of these populations to acidification has important implications for regional restoration efforts (e.g., Tettelbach et al., 2015). Future efforts in regions where both bivalve species exist and diurnal patterns of pH and DO are extreme may seek to focus on the more resilient, M. mercenaria, over A. irradians, which is more susceptible to the imminent intensification of acidification predicted to occur this century (Doney et al., 2009; O'Leary et al. submitted).

ABLE 5 Mean (± standard deviation) pH, dissolved oxygen (DO), pCO ₂ , saturation states of calcite and aragonite, total dissolved inorganic carbon
TDIC), carbonate, total alkalinity (TA), salinity, and temperature for the four treatment larval Argopecten irradians diurnal acidification and hypoxia
experiment.

Parameter	Continuous				Diurnal		
	Control	Intermediate	Low	Day	Night	Mean	
pH _T	7.89 ± 0.03	7.48 ± 0.05	7.22 ± 0.05	7.89 ± 0.06	7.28 ± 0.11	7.61 ± 0.26	
DO (mg L^{-1})	6.87 ± 0.25	4.08 ± 0.41	1.38 ± 0.45	7.09 ± 0.12	1.32 ± 0.49	4.11 ± 2.80	
pCO ₂ (µatm)	542 ± 55	1590 ± 236	3030 ± 213	530 ± 44	2820 ± 336	1670 ± 1200	
$\Omega_{calcite}$	2.83 ± 0.10	1.17 ± 0.08	0.67 ± 0.06	2.87 ± 0.18	0.68 ± 0.06	1.77 ± 1.10	
$\Omega_{aragonite}$	1.83 ± 0.06	0.76 ± 0.06	0.43 ± 0.04	1.86 ± 0.12	0.44 ± 0.04	1.15 ± 0.74	
TDIC (μ mol L ⁻¹)	1790 ± 58	1920 ± 136	2020 ± 45	1780 ± 45	1960 ± 21	1870 ± 100	
CO_3^{2-} (µmol L ⁻¹)	114 ± 3.6	47.0 ± 3.5	26.8 ± 2.6	115 ± 7.5	27.4 ± 2.3	71.3 ± 46	
TA (μ mol L ⁻¹)	1940 ± 56	1940 ± 131	1970 ± 46	1940 ± 49	1920 ± 11	1910 ± 35	
Salinity	30.2 ± 1.2						
Temperature (°C)	23.3 ± 0.84	23.3 ± 0.84	23.3 ± 0.84	23.3 ± 0.84	23.3 ± 0.84	23.3 ± 0.84	



FIGURE 7 | Survival (A), growth **(B)**, and development **(C)** of *Mercenaria mercenaria* larvae in the diurnal pH - DO experiment **(Table 6)**. Percent metamorphosis was determined 12 days post-fertilization. Error bars represent standard deviation (n = 4). Lowercase letters indicate significant differences among treatments.

A growing body of literature has documented the common and widespread nature of concurrent hypoxia and acidification in estuaries, particularly those that are eutrophic (Cai et al., 2011; Melzner et al., 2012; Wallace et al., 2014) and has shown that these conditions can be chronic or can vary diurnally in parallel with ecosystem metabolism (Ringwood and Keppler, 2002; Yates et al., 2007; Baumann et al., 2014). Due to the intimate linkage of DO and CO₂ associated with photosynthesis and respiration, the examination of only acidification or hypoxia without consideration of the other costressor is environmentally unrealistic. Similarly, in shallow, eutrophic estuaries, the consideration of only the continuous exposure of marine animals to acidification and hypoxia may not be reflective of ecosystem conditions. This study, therefore, sought to consider concurrent, diel changes in pH and DO reflective of the manner in which these variables cycle in shallow, temperate estuaries during summer (Ringwood and Keppler, 2002; Yates et al., 2007; Baumann et al., 2014). The patterns, intensity, and duration of diel acidification and hypoxia in coastal ecosystems can result in an infinite number of experimental combinations and conditions. By using two levels of acidification and hypoxia, this study expanded on prior investigations (Clark and Gobler, 2016) and provided two benchmark conditions to compare the diurnally cycling pH and DO to: the same mean conditions and the low, minimum condition achieved at night only. While the absolute levels of low pH and DO used in this study were more severe than coastal shelf ecosystems (Wootton et al., 2008; Hofmann et al., 2011; Cornwall et al., 2013) they were consistent with observations made in temperate estuaries during summer months (Ringwood and Keppler, 2002; Baumann et al., 2014).

During this study, the combined effects of hypoxia and acidification were more severe than those observed for acidification alone. For both larval clams and scallops, growth, and survival rates under diurnally-fluctuating pH and DO were significantly lower than the rates for individuals continuously exposed to the same mean pH and DO levels (e.g., intermediate conditions in **Figures 6**, 7). Instead, survival rates for both species under diurnally-fluctuating pH and DO were equal to rates experienced by individuals in the "low" treatments that were, on average, chronically exposed to pH levels 0.3-0.4 units lower and DO levels more than 2 mg L^{-1} lower. This

Parameter	Continuous			Diurnal		
	Control	Intermediate	Low	Day	Night	Mean
pH _T	7.87 ± 0.03	7.43 ± 0.03	7.24 ± 0.04	7.78 ± 0.06	7.24 ± 0.07	7.41 ± 0.34
DO (mg L^{-1})	6.92 ± 0.13	3.92 ± 0.34	1.32 ± 0.30	7.08 ± 0.15	1.34 ± 0.59	4.02 ± 3.00
pCO ₂ (µatm)	612 ± 36	1750 ± 142	2710 ± 203	694 ± 96	2850 ± 544	1770 ± 1170
$\Omega_{calcite}$	2.81 ± 0.17	1.23 ± 0.14	0.77 ± 0.14	2.41 ± 0.36	0.76 ± 0.17	1.58 ± 0.90
$\Omega_{aragonite}$	1.82 ± 0.11	0.79 ± 0.09	0.50 ± 0.04	1.56 ± 0.25	0.49 ± 0.11	1.02 ± 0.58
TDIC (μ mol L ⁻¹)	1880 ± 32	2060 ± 97	2050 ± 36	1840 ± 105	2050 ± 32	1940 ± 133
CO_3^{2-} (µmol L ⁻¹)	113 ± 6.9	49.2 ± 5.5	31.0 ± 2.2	96.6 ± 15	30.4 ± 6.9	63.5 ± 36
TA (μ mol L ⁻¹)	2030 ± 38	2080 ± 101	2010 ± 34	1960 ± 117	2010 ± 34	1980 ± 87
Salinity	30.3 ± 0.35	30.3 ± 0.35	30.3 ± 0.35	30.3 ± 0.35	30.3 ± 0.35	30.3 ± 0.35
Temperature (°C)	22.9 ± 0.16	22.9 ± 0.16	22.9 ± 0.16	22.9 ± 0.16	22.9 ± 0.16	22.9 ± 0.16

TABLE 6 | Mean (± standard deviation) pH, dissolved oxygen (DO), pCO₂, saturation states of calcite and aragonite, total dissolved inorganic carbon (TDIC), carbonate, total alkalinity (TA), salinity, and temperature for the four treatment larval *Mercenaria mercenaria* diurnal acidification and hypoxia experiment.

finding of diurnal fluctuations yielding more dire physiological outcomes is consistent with studies of isopods and calcifying algae exposed to diurnal acidification (Alenius and Munguia, 2012; Cornwall et al., 2013). It is also consistent with our prior study examining only one level of pH and DO during which the performance of larval bivalves experiencing diurnal acidification and/or hypoxia was generally worse than that of individuals exposed to the same mean conditions (Clark and Gobler, 2016). Here, by examining multiple levels of pH and DO, the quantitative extent of impairment due to diurnal changes in acidification and hypoxia was identified and affirmed that large diurnal changes in pH and DO have strong negative effects on early-life stage bivalves. While bivalves have a series of physiological adaptions to acclimate to low pH and low DO including acid-base regulation, anaerobic metabolism, and metabolic depression (Grieshaber et al., 1994; Guppy and Withers, 1999; Hochachka and Lutz, 2001; Wu, 2002; Michaelidis et al., 2005), it would seem these processes are ineffective against the combined level and variation of acidification and hypoxia at least for larval bivalves which may have more poorly developed physiological systems for resisting physiological stress. While it is probable that juvenile-stage bivalves with more well-developed physiological systems may be better able to adapt to diurnal changes on pH and DO, this has yet to be explored.

In conclusion, the effects of diurnal acidification alone on early-life stage bivalves were consistent with the effects consistent exposure to the same mean levels of pH, indicating the individuals could tolerate extreme acidification if normal pH conditions were encountered for an extended period daily. In contrast, the dual, diurnal variation in pH and DO associated with acidification and hypoxia yielded survival rates for larval bivalves that were not statistically different to individuals experiencing drastically lower levels of pH and DO. Given that high rates of ecosystem metabolism in eutrophic estuaries can causes pH and DO to undergo large diurnal changes (Ringwood and Keppler, 2002; Yates et al., 2007; Baumann et al., 2014) during the times of the year when early-life stage bivalves are present in estuaries (Kraeuter and Castagna, 2001; Shumway and Parsons, 2011), it is likely that these conditions have diminished and shaped abundances and growth rates in some bivalve populations (O'Leary et al. submitted). In the absence of nutrient abatement to lessen extremes in eutrophic-driven acidification and hypoxia (Wallace et al., 2014), it should be expected that climate change will lead to both greater extremes and lower levels of pH and DO (Rabalais et al., 2002; Miller et al., 2009; Feely et al., 2010) and thus lead to even greater losses for these populations.

ETHICS STATEMENT

All organisms used for this research were marine invertebrate bivalves and as such were not subject to Institutional Animal Care and Use Committee (IACUC; USA) oversight. However, care was taken in designing experiments to limit and reduce the number animals sacrificed during the course of research. All organisms prior to and after experimentation were adequately fed, maintained at optimal densities, and routinely cared for.

AUTHOR CONTRIBUTIONS

CG-obtained funding, designed experiments, analyzed data, and wrote manuscript. HC-designed experiments, conducted experiments, analyzed experimental results, and wrote manuscript. AG-designed experiments, analyzed experimental results, and wrote manuscript. ML-aided in conducting experiments and edited manuscript.

FUNDING

This work was supported by New York Sea Grant Award R-FBM-38, the Laurie Landeau Foundation, the Simons Foundation, and the Chicago Community Trust.

REFERENCES

- Abeels, H. A., Loh, A. N., and Volety, A. K. (2012). Trophic transfer and habitat use of oyster *Crassostrea virginica* reefs in southwest Florida, identified by stable isotope analysis. *Mar. Ecol. Prog. Ser.* 462, 125–142. doi: 10.3354/meps 09824
- Alenius, B., and Munguia, P. (2012). Effects of pH variability on the intertidal isopod, *Paradella dianae. Mar. Freshw. Behav. Physiol.* 45, 245–259. doi: 10.1080/10236244.2012.727235
- Baumann, H., Wallace, R. B., Tagliaferri, T., and Gobler, C. J. (2014). Large natural pH, CO₂ and O₂ fluctuations in a temperate tidal salt marsh on diel, seasonal, and interannual time scales. *Estuar. Coasts* 38, 220–231. doi: 10.1007/s12237-014-9800-y
- Caffrey, J. M. (2004). Factors controlling net ecosystem metabolism in U.S. estuaries. *Estuaries* 27, 90–101. doi: 10.1007/BF02803563
- Cai, W.-J., Hu, X., Huang, W.-J., Murrell, M. C., Lehrter, J. C., Lohrenz, S. E., et al. (2011). Acidification of subsurface coastal waters enhanced by eutrophication. *Nat. Geosci.* 4, 766–770. doi: 10.1038/ngeo1297
- Clark, H. R., and Gobler, C. J. (2016). Do diurnal fluctuations in CO₂ and dissolved oxygen concentrations provide a refuge from hypoxia and acidification for early life stage bivalves? *Mar. Ecol. Prog. Ser.* 558, 1–14. doi: 10.3354/meps11852
- Coen, L. D., Brumbaugh, R. D., Bushek, D., Grizzle, R., Luckenbach, M. W., Posey, M. H., et al. (2007). Ecosystem services related to oyster restoration. *Mar. Ecol. Prog. Ser.* 341, 303–307. doi: 10.3354/meps341303
- Cornwall, C. E., Hepburn, C. D., McGraw, C. M., Currie, K. I., Pilditch, C. A., Hunter, K. A., et al. (2013). Diurnal fluctuations in seawater pH influence the response of a calcifying macroalga to ocean acidification. *Proc. Biol. Sci.* 280:20132201. doi: 10.1098/rspb.2013.2201
- Del Giorgio, P. A., and Williams, P. J. (2005). *Respiration in Aquatic Ecosystems*. New York, NY: Oxford University Press.
- Diaz, R. J., and Rosenberg, R. (2008). Spreading dead zones and consequences for marine ecosystems. *Science* 321, 926–929. doi: 10.1126/science.1156401
- Dickson, A. G. (1993). The measurement of sea water pH. *Mar. Chem.* 44, 131–142. doi: 10.1016/0304-4203(93)90198-W
- Doney, S. C., Fabry, V. J., Feely, R. A., and Kleypas, J. A. (2009). Ocean acidification: the other CO₂ problem. Ann. Rev. Mar. Sci. 1, 169–192. doi: 10.1146/annurev.marine.010908.163834
- Doney, S. C., Ruckelshaus, M., Duffy, J. E., Barry, J. P., Chan, F., English, C. A., et al. (2012). Climate change impacts on marine ecosystems. *Ann. Rev. Mar. Sci.* 4, 11–37. doi: 10.1146/annurev-marine-041911-111611
- Ekstrom, J. A., Suatoni, L., Cooley, S. R., Pendleton, L. H., Waldbusser, G. G., Cinner, J. E., et al. (2015). Vulnerability and adaptation of US shellfisheries to ocean acidification. *Nat. Clim. Chang.* 5, 207–214. doi: 10.1038/nclimate2508
- Feely, R. A., Alin, S. R., Newton, J., Sabine, C. L., Warner, M., Devol, A., et al. (2010). The combined effects of ocean acidification, mixing, and respiration on pH and carbonate saturation in an urbanized estuary. *Estuar. Coast. Shelf Sci.* 88, 442–449. doi: 10.1016/j.ecss.2010.05.004
- Fegley, S. R. (2001). "Demography and dynamics of hard clam populations," in *Biology of the Hard Clam*, eds J. N. Kraeuter and M. Castagna (Amsterdam: Elsevier Science), 383–422.
- Gazeau, F., Parker, L. M., Comeau, S., Gattuso, J.-P., O'Connor, W. A., Martin, S., et al. (2013). Impacts of ocean acidification on marine shelled molluscs. *Mar. Biol.* 160, 2207–2245. doi: 10.1007/s00227-013-2219-3
- Gobler, C. J., and Baumann, H. (2016). Hypoxia and acidification in ocean ecosystems: coupled dynamics and effects on marine life. *Biol. Lett.* 12:20150976. doi: 10.1098/rsbl.2015.0976
- Gobler, C. J., DePasquale, E. L., Griffith, A. W., and Baumann, H. (2014). Hypoxia and acidification have additive and synergistic negative effects on the growth, survival, and metamorphosis of early life stage bivalves. *PLoS ONE* 9:e83648. doi: 10.1371/journal.pone.0083648
- Gobler, C. J., and Talmage, S. C. (2013). Short- and long-term consequences of larval stage exposure to constantly and ephemerally elevated carbon dioxide for marine bivalve populations. *Biogeosciences* 10, 2241–2253. doi: 10.5194/bg-10-2241-2013
- Green, M. A., and Aller, R. C. (1998). Seasonal patterns of carbonate diagenesis in nearshore terrigenous muds: relation to spring phytoplankton bloom and temperature. J. Mar. Res. 56, 1097–1123. doi: 10.1357/002224098765 173473

- Green, M. A., Waldbusser, G. G., Reilly, S. L., Emerson, K., and O'Donnell, S. (2009). Death by dissolution: sediment saturation state as a mortality factor for juvenile bivalves. *Limnol. Oceanogr.* 54, 1037–1047. doi: 10.4319/lo.2009.54.4.1037
- Grieshaber, M. K., Hardewig, I., Kreutzer, U., and Pörtner, H. O. (1994). Physiological and metabolic responses to hypoxia in invertebrates. *Rev. Physiol. Biochem. Pharmacol.* 125, 43–147. doi: 10.1007/bfb0030909
- Guppy, M., and Withers, P. (1999). Metabolic depression in animals: physiological perspectives and biochemical generalizations. *Biol. Rev. Camb. Philos. Soc.* 74, 1–40. doi: 10.1017/S0006323198005258
- Helm, M. M., Bourne, N., and Lovatelli, A. (2004). *Hatchery Culture of Bivalves:* A Practical Manual. Rome: Food and Agriculture Organization of the United Nations.
- Hochachka, P. W., and Lutz, P. L. (2001). Mechanism, origin, and evolution of anoxia tolerance in animals. *Comp. Biochem. Physiol. B Biochem. Mol. Biol.* 130, 435–459. doi: 10.1016/S1096-4959(01)00408-0
- Hofmann, G. E., Smith, J. E., Johnson, K. S., Send, U., Levin, L. A., Micheli, F., et al. (2011). High-frequency dynamics of ocean pH: a multi-ecosystem comparison. *PLoS ONE* 6:e28983. doi: 10.1371/journal.pone.0028983
- Keppel, A. G., Breitburg, D. L., and Burrell, R. B. (2016). Effects of co-varying dielcycling hypoxia and pH on growth in the Juvenile Eastern Oyster, *Crassostrea* virginica. PLoS ONE 11:e0161088. doi: 10.1371/journal.pone.0161088
- Kim, T. W., Barry, J. P., and Micheli, F. (2013). The effects of intermittent exposure to low-pH and low-oxygen conditions on survival and growth of juvenile red abalone. *Biogeosciences* 10, 7255–7262. doi: 10.5194/bg-10-7255-2013
- Kraeuter, J. N., and Castagna, M. (2001). *Biology of the Hard Clam.* Amsterdam: Elsevier.
- Kraeuter, J. N., Castagna, M., and van Dessel, R. (1981). Egg size and larval survival of *Mercenaria mercenaria* (L.) and *Argopecten irradians* (Lamarck). *J. Exp. Mar. Biol. Ecol.* 56, 3–8. doi: 10.1016/0022-0981(81)90003-4
- Kroeker, K. J., Kordas, R. L., Crim, R., Hendriks, I. E., Ramajo, L., Singh, G. S., et al. (2013). Impacts of ocean acidification on marine organisms: quantifying sensitivities and interaction with warming. *Glob. Chang. Biol.* 19, 1884–1896. doi: 10.1111/gcb.12179
- Melzner, F., Thomsen, J., Koeve, W., Oschlies, A., Gutowska, M. A., Bange, H. W., et al. (2012). Future ocean acidification will be amplified by hypoxia in coastal habitats. *Mar. Biol.* 160, 1875–1888. doi: 10.1007/s00227-012-1954-1
- Meyer, D. L., Townsend, E. C., and Thayer, G. W. (1997). Stabilization and erosion control value of oyster cultch for intertidal marsh. *Restor. Ecol.* 5, 93–99. doi: 10.1046/j.1526-100X.1997.09710.x
- Michaelidis, B., Ouzounis, C., Paleras, A., and Pörtner, H. O. (2005). Effects of long-term moderate hypercapnia on acid-base balance and growth rate in marine mussels *Mytilus galloprovincialis*. *Mar. Ecol. Prog. Ser.* 293, 109–118. doi: 10.3354/meps293109
- Miller, A. W., Reynolds, A. C., Sobrino, C., and Riedel, G. F. (2009). Shellfish face uncertain future in high CO₂ world: influence of acidification on oyster larvae calcification and growth in estuaries. *PLoS ONE* 4:e5661. doi: 10.1371/journal.pone.0005661
- Millero, F. J., Graham, T. B., Huang, F., Bustos-Serrano, H., and Pierrot, D. (2006). Dissociation constants of carbonic acid in seawater as a function of salinity and temperature. *Mar. Chem.* 100, 80–94. doi: 10.1016/j.marchem.2005.12.001
- Officer, C. B., Smayda, T. J., and Mann, R. (1982). Benthic filter feeding a natural eutrophication control. *Mar. Ecol. Prog. Ser.* 9, 203–210. doi: 10.3354/meps009203
- Paulmier, A., Ruiz-Pino, D., and Garçon, V. (2011). CO₂ maximum in the oxygen minimum zone (OMZ). *Biogeosciences* 8, 239–252. doi: 10.5194/bg-8-239-2011
- Pörtner, H. (2008). Ecosystem effects of ocean acidification in times of ocean warming: a physiologist's view. *Mar. Ecol. Prog. Ser.* 373, 203–217. doi: 10.3354/meps07768
- Pörtner, H. O., Langenbuch, M., and Reipschlager, A. (2004). Biological impact of elevated ocean CO₂ concentrations: lessons from animal physiology and earth history. J. Oceanogr. 60, 705–718. doi: 10.1007/s10872-004-5763-0
- Rabalais, N. N., Turner, R. E., and Wiseman, W. J. (2002). Gulf of Mexico Hypoxia, A.K.A. "The Dead Zone". Annu. Rev. Ecol. Syst. 33, 235–263. doi: 10.1146/annurev.ecolsys.33.010802.150513
- Ringwood, A. H., and Keppler, C. J. (2002). Water quality variation and clam growth: is pH really a non-issue in estuaries? *Estuaries* 25, 901–907. doi: 10.1007/BF02691338

- Shumway, S. E., Davis, C., Downey, R., Karney, R., Kraeuter, J., Parsons, J., et al. (2003). Shellfish aquaculture-in praise of sustainable economies and environments. *World Aquac.* 34, 8–10.
- Shumway, S. E., and Parsons, G. J. (2011). Scallops: Biology, Ecology and Aquaculture. Amsterdam: Elsevier.
- Sokolova, I. M. (2013). Energy-limited tolerance to stress as a conceptual framework to integrate the effects of multiple stressors. *Integr. Comp. Biol.* 53, 597–608. doi: 10.1093/icb/ict028
- Steckbauer, A., Ramajo, L., Hendriks, I. E., Fernandez, M., Lagos, N. A., Prado, L., et al. (2015). Synergistic effects of hypoxia and increasing CO₂ on benthic invertebrates of the central Chilean coast. *Front. Mar. Sci.* 2:49. doi: 10.3389/fmars.2015.00049
- Talmage, S. C., and Gobler, C. J. (2009). The effects of elevated carbon dioxide concentrations on the metamorphosis, size, and survival of larval hard clams (*Mercenaria mercenaria*), bay scallops (*Argopecten irradians*), and Eastern oysters (*Crassostrea virginica*). *Limnol. Oceanogr.* 54, 2072–2080. doi: 10.4319/lo.2009.54.6.2072
- Talmage, S. C., and Gobler, C. J. (2010). Effects of past, present, and future ocean carbon dioxide concentrations on the growth and survival of larval shellfish. *Proc. Natl. Acad. Sci. U.S.A.* 107, 17246–17251. doi: 10.1073/pnas.0913804107
- Talmage, S. C., and Gobler, C. J. (2011). Effects of elevated temperature and carbon dioxide on the growth and survival of larvae and juveniles of three species of northwest Atlantic bivalves. *PLoS ONE* 6:e26941. doi: 10.1371/journal.pone.0026941
- Tettelbach, S. T., Peterson, B. J., Carroll, J. M., Furman, B. T., Hughes, S. W. T., Havelin, J., et al. (2015). Aspiring to an altered stable state: rebuilding of bay scallop populations and fisheries following intensive restoration. *Mar. Ecol. Prog. Ser.* 529, 121–136. doi: 10.3354/meps11263
- Tolley, S. G., and Volety, A. K. (2005). The role of oysters in habitat use of oyster reefs by resident fishes and decapod crustaceans. *J. Shellfish Res.* 24, 1007–1012. doi: 10.2983/0730-8000(2005)24[1007:TROOIH]2.0.CO;2
- Vance, R. R. (1973). On reproductive strategies in marine benthic invertebrates. *Am. Nat.* 107, 339–352. doi: 10.1086/282838
- Waldbusser, G. G., Brunner, E. L., Haley, B. A., Hales, B., Langdon, C. J., and Prahl, F. G. (2013). A developmental and energetic basis linking larval oyster shell formation to acidification sensitivity. *Geophys. Res. Lett.* 40, 2171–2176. doi: 10.1002/grl.50449

- Waldbusser, G. G., Hales, B., Langdon, C. J., Haley, B. A., Schrader, P., Brunner, E. L., et al. (2014). Saturation-state sensitivity of marine bivalve larvae to ocean acidification. *Nat. Clim. Chang.* 5, 273–280. doi: 10.1038/nclimate 2479
- Waldbusser, G. G., Hales, B., Langdon, C. J., Haley, B. A., Schrader, P., Brunner, E. L., et al. (2015). Ocean acidification has multiple modes of action on bivalve larvae. *PLoS ONE* 10:e0128376. doi: 10.1371/journal.pone.01 28376
- Wall, C. C., Peterson, B. J., and Gobler, C. J. (2008). Facilitation of seagrass Zostera marina productivity by suspension-feeding bivalves. Mar. Ecol. Prog. Ser. 357, 165–174. doi: 10.3354/meps07289
- Wallace, R. B., Baumann, H., Grear, J. S., Aller, R. C., and Gobler, C. J. (2014). Coastal ocean acidification: the other eutrophication problem. *Estuar. Coast. Shelf Sci.* 148, 1–13. doi: 10.1016/j.ecss.2014.05.027
- Widdicombe, S., and Spicer, J. I. (2008). Predicting the impact of ocean acidification on benthic biodiversity: what can animal physiology tell us? J. Exp. Mar. Biol. Ecol. 366, 187–197. doi: 10.1016/j.jembe.2008.07.024
- Wootton, J. T., Pfister, C. A., and Forester, J. D. (2008). Dynamic patterns and ecological impacts of declining ocean pH in a high-resolution multi-year dataset. *Proc. Natl. Acad. Sci. U.S.A.* 105, 18848–18853. doi:10.1073/pnas.0810079105
- Wu, R. S. (2002). Hypoxia: from molecular responses to ecosystem responses. Mar. Pollut. Bull. 45, 35–45. doi: 10.1016/S0025-326X(02)00061-9
- Yates, K. K., Dufore, C., Smiley, N., Jackson, C., and Halley, R. B. (2007). Diurnal variation of oxygen and carbonate system parameters in Tampa Bay and Florida Bay. *Mar. Chem.* 104, 110–124. doi: 10.1016/j.marchem.2006.12.008

Conflict of Interest Statement: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

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